Assessment of Radioactivity contents and associated risks in some soil used for agriculture and building materials in Cameroon

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Abstract

A survey of radioactivity content and associated radiological risks was carried out in various soils used for agriculture and building materials in Cameroon by means of a well-calibrated high-purity germanium detector. Soil samples were collected directly from the agricultural farms and the brick's factories, air-dried at room temperature to a constant mass, crushed, sieved and sealed for at least one month before analysis. The specific activity of ^{238}U ranged from 5.36 ± 0.39 to 51.28 ± 9.67 Bq kg^{-1} with an average of 34.52 ± 7.18 Bq kg^{-1}; ^{232}Th from 4.03 ± 1.03 to 24.74 ± 3.10 Bq kg^{-1} with an average of 16.67 ± 4.28 Bq kg^{-1}; ^{40}K ranged from 16.18 ± 3.11 to 467.40 ± 50.80 Bq kg^{-1} with an average of 186.96 ± 16.21 Bq kg^{-1}; while that of the fallout ^{137}Cs ranged from 0.00 to 4.79 ± 1.75 Bq Kg^{-1} with an average of 2.32 ± 0.99 Bq kg^{-1}; The mean result obtained for the Representative levels index (I_γ), the radium equivalent (Ra_{eq}), the total absorbed dose rate (ADR), the external hazard index (H_{ex}), the internal hazard index (H_{in}) and the Ceasium intervention levels were 0.52, 72.75 Bq kg^{-1}, 33.73 nGy h^{-1}, 0.20 Bq kg^{-1}, 0.29 Bq kg^{-1} and 0.0028 Bq cm^2 respectively. The discrepancies of our data can be attributed to several factors such as past nuclear disasters, geological formation, transport process, etc. Although our results are just some fractions of the international standard limit, but still within the same ranges when compared with those obtained elsewhere. This results also will serve as a baseline data for future investigations.

Keywords: Radioactivity, soil, hazards, Gamma spectroscopy, Cameroon

1.   Introduction

Radionulides that accumulate in soils constitute a direct route of exposure to human population when the contaminated soils are used. Thus, external exposure of human to ionizing radiation from the use of soil has been noted with great concern (Abbady 2004, Kirchner et al. 2002; Bikit et al. 2001; Mollah et al. 1986a; Farai and Ademola 2001; Olomo et al. 2003; Mollah et al. 1986b). Soil is of great important in human life. It constitutes the ground from which man settlement is build and through which foods are produced. Soils and rocks contain natural radionuclides of a wide range of activity levels. This radioactivity can be transported to the surface by several processes: tectonic movement, volcanic activity, ore/coal mining and groundwater flow. The transportation by groundwater flow and similar mechanism depends on the solubility and chemical reactions of the radioactive substance, and the porosity of the ground. Migration plays an important role in the final storage of radioactive waste. The most severe civil nuclear reactor accident at Chernobyl power station (IAEA 1986; Belayev et al. 1990) occurred on 26th April 1986. The post-impact assessment program for the estimation of migration dose as a component of radiation burden of the populace in Cameroun was put in place because of the worldwide distribution of its radioactive effluents.

Cameroon is surrounded by countries having nuclear reactors and it is not far from the Sahara where some atmospheric weapons tests took place (Figure 1). With particular reference of the lessons and experiences of the Chernobyl-4 accident, there is a growing need to assure the Cameroon public over the concern about the possibility of nuclear accident with transboundary consequences which may nevertheless occur in spite of any adequate nuclear safety criteria by countries operating nuclear(research or power) reactors, particularly in Africa. During the years after those nuclear accidents, the bioactivity

and environmental mobility of radiocesium declines markedly, resulting in large changes in contamination of soils, surface water, foodstuffs, etc. The sorption of radiocesium to soils and sediments has shown that Cs is selectively bound to specific sorption sites on illitic clay minerals where it is available for ion-exchange with ions, which have a similar hydrated radius, specifically potassium and ammonium (Smith et al. 1999). Over time, however radiocesium slowly diffuses into the illite lattice know as "fixation". Radiocesium fixation in soils and sediments are crucial due to its long-term mobility and bioavailability.

In recent times, there is an increased awareness of the effects of environmental radiation from natural radioactivity which has been enhanced by some environmental fields such as rocks, soils, water air, etc. Attention has been given to radioactivity contents and radiation levels in various types of solid due to naturally occurring radioactivity. Radiation, by its very nature is harmful to life. Depending on the doses, it can induce cancer, genetic and organs damages, cell killing and also rapid death.

Many research works have been carried out on radioactivity measurement throughout the world (Abbady 2004; Bikit et al. 2001; Mollah et al. 1986b; Papp et al. 2002). Here in Nigeria, Tchokossa and his co-workers reported the results of many investigations on the distribution of radioactivity content in various environmental matrices (Tchokossa et al. 1999; Fasasi et al. 1999; Fasasi et al. 2003; Avwiri and Tchokossa 2006; Tchokossa 2006; Avwiri et al. 2007; Awodugba and Tchokossa 2008; Ajayi et al. 2009), while in Cameroon, Makon et al. carried out similar research in some edible vegetable cultivated and consumed worldwide(Makon et al. 2011). Cameroun is located at the center of both operational and non-operational nuclear reactors and industries(Figure 1), and to the best of my knowledge, there is no published or on-going works on the survey of Cameroonian's soil.

This present work aimed to provide information on the radiation levels in the soil and the resulting contribution of past nuclear operation carried out that may have being accumulated in soil. The data generated will not only help to design a radiation mapping for agricultural lands and building materials, but also will be use as baseline data for future investigations, government, policy makers and the general public.

2.   Materials and Methods

A total of 45 soil samples were collected from nine towns across five provinces of the country. At each location, well described elsewhere(Makon et al., 2011), the samples were collected at the depth between 0 – 30 cm, at the four corners and the center of a square frame of 1 m size and later composed to make a single sample. Samples were collected in their natural form, dried at room temperature to a constant weight, grounded and sieved to a 2 mm size. Each sample was sealed in cylindrical plastic containers. The standard reference sample and the background which was an empty container have the same geometry like the sample containers. The sealed samples were kept for at least 28 days to allow a stage of secular equilibrium to reach before gamma spectrometry analysis commenced.

The counting equipment well described elsewhere(Tchokossa et al. 1999; Tchokossa 2006) was a Canberra vertical high – purity coaxial germanium (HPGe) crystal detector, model GC 2018-7500, serial number b87063 enclosed in a 100 mm thick lead shield and coupled to a Canberra Multichannel Analysing (MCA) computer system. The energy and efficiency calibrations were done using a well calibrated standard soils sources supplied by the International Atomic Energy Agency (IAEA), Vienna, Austria (Reference Material IAEA – 375 for radionuclides and trace elements in soil) and some radioactive point sources from the Isotopes Products Laboratories, Burbank, California, USA. The advantage of gamma ray spectrometry is its ability to measure gamma emitters directly in the original sample without the need for chemical separation. It allows both qualitative identification and quantitative determination of the radionuclides in the sample. The counting time for both the sample and the background was 36,000 s. The photopeaks observed with regularity were identified to belong not only to the naturally occurring series decay radionuclide headed by ^{238}U and ^{232}Th, and non-series natural radionuclide ^{40}K, but also the presence of ^{137}Cs in most of the soils samples investigated. Other

radionuclides if present, appeared rather infrequently at low levels or occurred at levels below the maximum detectable limits (MDL) statistically determined a two-standard deviation analytical error. The characteristics of the selected gamma-emitting radionuclides used as monitor lines for this work are as shown in table 1. The ^{226}Ra (^{238}U) activity was estimated from ^{214}Pb at 351.9 keV, ^{214}Bi at 609.3 and 1120.3 keV;   that of ^{232}Th was from ^{208}Tl at 583.2 keV and ^{228}Ac at 911.1 keV. ^{40}K and ^{137}Cs were determined directly from the 1460.3 and 661.6 keV gamma-emission respectively. The activities of the radionuclides in the measured samples were computed from the difference between the net peak and net background areas, accumulation time, absolute peak efficiency, absolute γ-ray emission probability and the sample mass. Duplicate analyses were conducted on all soil samples to check for the reproductivity of the results and the stability of the counting system. The total uncertainty in the measured concentrations was estimated from the combined uncertainties in the homogeneity of the samples, their position with respect to the detector, the calibration procedure, the peak area determination and the background correction.

3.   Results and Discussion

3.1   Activity Concentration

The average concentrations of the radionuclides observed with regularity in the soil samples analyzed by gamma-spectrometry are given in Table 2.

Table 2: Activity Concentration (Bq kg^{-1})

Locations	^{40}K		^{238}U		^{232}Th		^{137}Cs	
	Range	**Mean**	**Range**	**Mean**	**Range**	**Mean**	**Range**	**Mean**
1	177.58 – 265.02	221.30 ± 16.60	32.79 – 42.82	37.61 ± 6.03	17.76 – 23.40	21.45 ± 5.63	3.96 – 4.31	4.14 ± 1.38
2	66.19 – 292.65	156.06 ± 12.14	27.57 – 51.28	39.34 ± 9.38	10.62 – 20.04	16.11 ± 4.21	2.55 – 3.38	2.98 ± 1.38
3	32.58 – 122.07	76.48 ± 14.32	24.67 – 36.49	27.28 ± 6.69	10.47 – 18.34	15.18 ± 4.56	ND	ND
4	45.38 – 224.23	145.60 ± 10.92	5.36 – 42.85	26.73 ± 5.26	6.08 – 24.74	16.16 ± 2.30	1.32 – 4.79	3.06 ± 1.19
5	453.20 – 467.40	465.30 ± 43.70	46.92 – 49.71	49.22 ± 12.11	20.44 – 23.93	22.19 ± 7.37	ND – 2.87	2.66 ± 1.09
6	187.66 – 209.73	191.97 ± 14.78	32.40 – 49.38	40.89 ± 7.70	17.72 – 21.44	19.58 ± 6.06	ND – 2.24	2.08 ± 0.76
7	162.87 – 213.16	182.09 ± 13.83	39.86 – 49.86	47.13 ± 6.70	12.04 – 16.23	14.75 ± 3.25	ND – 2.04	1.90 ± 0.68
8	211.71 – 232.69	224.63 ± 17.47	27.30 – 38.24	34.46 ± 10.10	17.50 – 21.33	19.55 ± 3.86	ND – 2.32	2.28 ± 0.83

9	16.18 – 32.42	19.23 ± 2.14	6.80 – 10.22	8.00 ± 0.67	4.03 – 8.22	5.09 ± 1.27	ND – 1.98	1.76 ± 0.61
Mean	**16.18 – 467.40**	**186.96 ± 16.21**	**5.36 – 51.28**	**34.52 ± 7.18**	**4.03 – 24.74**	**16.67 ± 4.28**	**ND – 4.79**	**2.32 ± 0.99**

Except for ^{40}K and ^{137}Cs, all the gamma-lines detected came from the ^{238}U and ^{232}Th decay series. The total specific activity from each radionuclide determined at two-standard deviation analytical error was calculated on the basis of a mean value obtained from each sampling location taken from a particular depth. The ^{238}U concentrations ranged from 5.36 ± 0.39 to 51.28 ± 9.67 Bq kg^{-1} with an average of 34.52 ± 7.18 Bq kg^{-1}. ^{232}Th from 4.03 ± 1.03 to 24.74 ± 3.10 Bq kg^{-1} with an average of 16.67 ± 4.28 Bq kg^{-1}. ^{40}K from 16.18 ± 3.11 to 467.40 ± 50.80 Bq kg^{-1} with an average of 186.96 ± 16.21 Bq kg^{-1} and lastly ^{137}Cs from 0.00 to 4.79 ± 1.75 Bq kg^{-1} with an average 2.32 ± 0.99 Bq kg^{-1}.

For ^{40}K, the highest activity concentration was detected in location 5 (Bamenda), in fact geographically, the town is made up of series of mountains and valleys where most of the agricultural lands lie. The lowest was detected in location 9 (Limbe), probably because the sampling point is located to the upper slope part of the city, and the soil there are prone to a high level of erosion which may result in leaching of the soil. This pattern has been confirmed as we moved down the slope. This is favourable as a result of erosion which washes out soil to settle at the valley. ^{40}K is a naturally occurring radionuclide present abundantly in the earth crust as one of its major constituents. ^{40}K content in soil shows a more homogeneous distribution in the horizontal micro scale. The variability of ^{40}K among locations could be due to the differences in land management, for example, fertilization, plowing practices, geology and geography.

For ^{238}U, the highest activity concentration was found in location 2 (Ezeka) while the least was obtained in location 4 (Edea). In nature, uranium consists of primary two isotopes ^{238}U and ^{235}U, of which the former is the most abundant 99.73% (IAEA, 1979).

While for ^{232}Th, the highest activity concentration was obtained in Edea (location 4) and the least in Limbe (location 9). ^{238}U and ^{232}Th were detectable in all the soils sample with the former almost double that of ^{232}Th except in some few cases.

Geologically, the bedrocks consist of coarse to medium sand, subordinate slit, gneiss and granites which have affinity to ^{238}U and ^{232}Th. Slight variation in the radioactivity content in soil may be observed with different locations mainly due to the type of soil, soil formation and soil transport process.

For ^{137}Cs, the highest activity concentration was found in location 4 (Edea) while it was not detectable in location 3 (Douala). This higher presence of ^{137}Cs in locations 7, 1 and especially 4 may be attributed to the natural reactor in Oklo. In fact, during 1972 a natural nuclear reactor has been detected at Oklo in Gabon (South of Cameroun) (Bryant et al. 1976). It was critical about 1.8×10^9 years ago, and has generated totally 15,000 MW-years during his life-time of 600,000 years. The exact analysis of the composition of the fission products revealed that it was a light-water-moderated and- cooled reactor of the BWR (Boiling Water) type. The steam penetrated the overlaying rocks, dissolved further uranium ore, which was carried into the core by the condensate flow (automatic recharging during operation). The results of this natural long-term waste storage experiment can be compared to shorter man-made experiments of the order of 10 years duration (Hübschman 1988). It was pointed out that this waste was not encapsulated and that the main part of the plutonium is still in the reaction zone. Also Bamenda and Buea at locations 5 and 8 respectively were not too far from another natural disaster namely Lake Nyos.

Lake Nyos which is a crater-lake was formed a few hundred years ago in the highlands of Northwestern Cameroun during one of the volcanic eruptive activity. The Lake, located about 6° 26'N latitude and 10° 16'E longitude is 208 m deep and contains 1.32×10^8 m^3 of water (Lockwood et al., 1988). Geologically, the area lies within a volcanic active basement region with volcanic ask material

overlying the basement. The lake Nyos Dam was formed at the time of the formation of the Crater Lake. This Dam has been eroding at a geologically alarming rate. Granitic bedrock had therefore been exposed along a 40 m thick dam. This was left at its narrowest point where water overflows and spills from the lake through a surface path down the cliff-face of the dam during the rainy season. A lethal explosion led to the release of gas from Lake Nyos in the night of 21st August 1986, killing 1746 people, several thousand of cattle and other animals by suffocation as the gas cloud swept down the valleys to the North and through the valleys of Cha, Nyos and Subum. The gas explosion generated a huge seiche estimated as high as 75 m (Freeth et al., 1987) which uprooted trees and bamboo forest on the foreshore to the South of the lake. By September 14, 1986, the lake surface was still reddish brown in colour and littered with floating trees and debris. The disaster affected all settlements downstream within 25 km from the Lake. The gas explosion occurred without any previous warming, although two years earlier, on the 15th August 1984, a similar disastrous emission of carbon dioxide occurred in Lake Monoun, killing 37 people (Joint Team, 1991).

The occurrence of this fallout ^{137}Cs in soil in Cameroon may also be due to the Sahara's atmospheric nuclear weapon tests of the 50's and 60's which were the primary causes of high levels of global fallout. Radioactive materials were diffused depending on the extent and form of the emission, and are formed even in areas otherwise uninfluenced by human activity. It is known that small quantities of fission-products residues (^{137}Cs and ^{90}Sr) from atmospheric weapons test (AWTs) are slowly being purged from the food chain (Olomo 1990).

3.2 Estimation of the Radiological Hazard Indices

Table 3 present some radiological effects indices such as the Absorbed dose rate, the Radium Equivalent activity concentration, the Representative Level index, External Hazard Index (H_{ext}), Internal Hazard Index (H_{in}) and the Caesium -137 Intervention Levels.

3.2.1 Total Air Absorbed Dose Rate(D).

The total Absorbed Dose Rate D (in nGy hr^{-1}) at 1.0 m above the ground was estimated using the empirical formula: (Beck et al., 1972; UNSCEAR 1988).

$$D = 0.042A_K + 0.428A_U + 0.666A_{Th}$$

as a partial evaluation of the radiological hazard of ^{40}K, ^{238}U and ^{232}Th. Here A_K, A_U and A_{Th} are the specific activity concentration for ^{40}K, ^{238}U and ^{232}Th respectively. The results are contained in Table 3. The absorbed dose rate values ranged between 7.62 – 55.39 nGy hr^{-1} with a mean of 33.73 ± 6.46 nGy hr^{-1}. The least absorbed dose rate was recorded in location 9 (Limbe), while the highest was obtained in location 5 (Bamenda). Using an outdoor occupancy factor of 0.20 and the conversion factor of 0.70 Sv/Gy (UNSCEAR, 1988), the effective dose equivalent (H_E) arising from this average absorbed dose in air is 35.10 μSv/y. This value is just about 3.5% of the 1.0 mSv/y recommended by the International Commission on Radiological Protection (ICRP, 1991; NEA-OCED, 2011) as the maximum permissible H_E for members of the public. Therefore the area may be considered as safe for habitation.

3.2.2 Radium Equivalent Activity(Ra$_{eq}$)

In comparing the radioactivity of materials that contain Ra, Th and K, a common index termed Radium Equivalent activity (Ra$_{eq}$) is required to obtain the total activity and is also used to assess the gamma radiation hazards. Since 98% of the radiological effects of the uranium series are produced by radium and its daughter products (Sam and Abbass, 2001), the contribution from ^{238}U and the other ^{226}Ra precursors is usually ignored, so that the Raeq can be expressed as (Krieger, 1981; Beretka and Mathew 1985):

$$\text{Ra}_{eq} = 370 \text{ Bq/Kg} \left(\frac{A_{Ra}}{370.Bq/kg} + \frac{A_{Th}}{259Bq/kg} + \frac{A_K}{4810Bq/kg} \right)$$

Or

$$\text{Ra}_{eq} = A_{Ra} + 1.43 A_{Th} + 0.077 A_K$$

Where A_{Ra}, A_{Th} and A_K are the activity concentrations of ^{226}Ra, ^{232}Th and ^{40}K in Bq kg^{-1} respectively. The results of Ra$_{eq}$ ranged between 16.76 – 116.78 Bq kg^{-1} obtained in locations 9 (Limbe) and 5 (Bamenda) respectively, with an average of 72.75 Bq/kg. On the average, Raeq concentration was found to be 85.32 Bq kg^{-1} for location 1(Yaounde, the country's capital), 54.88 Bq kg^{-1} for location 3 (Douala, the most popular city). However, all the values obtained here for radium equivalent activity fall far below the criterion limit of 370 Bq kg^{-1}(Beretka and Mathew, 1985). It is apparent that the Ra$_{eq}$ originating from different regions shows some variations, which are likely to be related to the type of soil, transport process, etc. This is important in selecting suitable soils not only for construction but also for agricultural purposes.

3.2.3 Representative Level Index (I_γ)

Another radiation hazard index called The Representative Levels index I_{yr} is given by(NEA-OECD, 1979):

$$I_\gamma = \left(\frac{A_{Ra}}{150.Bq/kg} + \frac{A_{Th}}{100Bq/kg} + \frac{A_K}{1500Bq/kg} \right)$$

Where A_{Ra}, A_{Th} and A_K are the radioactivity concentrations of ^{226}Ra(or ^{238}U), ^{232}Th and ^{40}K in Bq/Kg respectively. The Representative Level Index ranged between 0.12 – 0.86 with an average of 0.52. The least and the highest were investigated in locations 9 (Limbe) and 5 (Bamenda). It should be noted that all I_γ values observed fall within a very narrow range and are less than unity, the upper limit for the representative level index. These confirm that the soil investigated exhibit very low gamma-radiation level.

3.2.4 External Hazard Index (H_{ext})

This index enable us to estimate the radiation dose expected to be delivered externally if a building is constructed using these materials. The external hazard index (H_{ex}) was obtained as follow (Saito et al. 1998, UNSCEAR, 2000):

$$H_{ex} = \left(\frac{A_{Ra}}{370} + \frac{A_{Th}}{259} + \frac{A_K}{4810} \right)$$

The results shown in table revealed that the external hazard index H_{ex} ranged between 0.05 and 0.31 Bq kg^{-1} from locations 9 (Limbe) and 5 (Bamenda) respectively, with a mean of 0.20 Bq kg^{-1}. The overall results still below the limit of 1 Bq kg^{-1}. Hence it is safe as building materials.

3.2.5 Internal hazard index (H_{in})

The internal exposure to ^{222}Rn and its progenies is controlled by the internal hazard index (H_{in}) which is given by (Quindos et al., 1987; Cottens 1990):

$$H_{in} = \left(\frac{A_{Ra}}{185} + \frac{A_{Th}}{259} + \frac{A_K}{4810} \right)$$

H_{in} ranged between 0.07 and 0.45 Bq kg^{-1} from locations 9 (Limbe) and 5 (Bamenda) with a mean of 0.29 Bq kg^{-1} as contained in table 1. Both H_{ex} and H_{in} recorded the lowest and the highest of their values in the same locations respectively. The highest value of H_{in} is still below the limit of 1 Bq kg^{-1} (Iqbal et al., 2000) for the safe use of material in the construction of dwellings.

3.2.6 Caesium Intervention (I_{Cs-137})

Also the Caesium-137 intervention was computed from the following equation (Walling and Quine 1993):

$$I_{Cs-137} = \left(\frac{A_{Cs-137}x.M}{A}\right)$$

Where A_{Cs-137} is the activity per unit mass of Cs-137

M: the mass of the sample converted in Kg and

A: the sampling area in cm^2.

From the table 1, I_{Cs-137} ranged between 0.0010 and 0.0071 Bq cm^{-2}, with a mean of 0.0028 Bq cm^{-2}. The lowest and the highest were obtained respectively in locations 7 (Lolodorf) and 9 (Limbe), due to the erosion process and underground migration of radionuclides. In fact Limbe is almost located at the sea levels.

4   Conclusion and Recommendations

Natural Occurring Radioactive Materials and man-made Cs-137 as well as the derived radiological hazards have being investigated in soil used for agriculture and building materials in some regions in Cameroon, using a well calibrated gamma-spectrometer system. The mean specific activity obtained was 186.96 ± 16.21, 34.52 ± 7.18, 16.67 ± 4.28 and 2.32 ± 0.99 Bq kg for ^{40}K, ^{238}U, ^{232}Th and ^{137}Cs respectively. The mean for the Absorbed Dose rate, Radium Equivalent, Representative Level Index, External hazard index, Internal hazard index and Ceasium Intervention Level were 3.73 nGy/h, 72.75 Bq/kg, 0.52, 0.20 Bq/kg, 0.29 Bq/kg and 0.0028 Bq/cm^2 respectively. These fall within the world ranges for building materials and soil. The relative higher radioactivity content in Bamenda could be attributed to the Athmospheric Nuclear weapon test and the natural disaster that took place in the 60's and 1986 respectively. Calculations of the derived radiological indices for all the investigated soils showed that none exceeds the recommended levels for external radiation exposure. In regard to the above results, the commonly used soil in those areas could be safely used for agriculture and building materials. Meanwhile better characterization of humus, soil mineralogy and regular radiological monitoring are necessary since accumulated doses may result in deleterious effects.

Table 3:   Derived Radiological Hazard Indices

Locations	Absorbed Dose Rate (nGy/hr)	Radium Equivalent Activity (Bq/kg)	Representative Levels Index I_γ	External Hazard Index (H_{ex}) (Bq/kg)	Internal Hazard Index (H_{in}) (Bq/kg)	Ceasium Intervention Levels (Bq/cm^2)
1	39.68	85.32	0.61	0.23	0.33	0.0021
2	34.12	74.39	0.53	0.20	0.31	0.0016
3	25.00	54.88	0.38	0.15	0.22	
4	28.32	61.05	0.44	0.16	0.24	0.0016
5	55.39	116.78	0.86	0.31	0.45	0.0012

6	38.60	83.67	0.60	0.23	0.34	0.0011
7	37.64	82.24	0.58	0.22	0.35	0.001
8	37.20	79.71	0.57	0.21	0.31	0.0068
9	7.62	16.76	0.12	0.05	0.07	0.0071
Mean	33.73	72.75	0.52	0.20	0.29	0.0028

Table 1: Characteristics of selected gamma-emitting radionuclides (IAEA, 1989)

Radionuclides	Source	Half-life $T_{1/2}$ (yr)	Gamma Energy E_γ(keV)	Gamma Intensity I_γ(%)
^{214}Pb	^{238}U decay	1.6×10^3	351.9	38.9
^{214}Bi	^{238}U decay	1.6×10^3	609.3	43.3
^{214}Bi	^{238}U decay	1.6×10^3	1120.3	15.7
^{238}Ac	^{232}Th decay	1.41×10^{10}	911.1	27.7
^{208}Tl	^{232}Th decay	1.41×10^{10}	583.2	85.8
^{40}K	primordial	1.28×10^9	1460.8	10.7
^{137}Cs	Chernobyl and Nuclear weapons tests	30	661.6	85.2

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